



El Colegio de la Frontera Sur

Degradación de plaguicidas por la microbiota intestinal de
abejas silvestres del Soconusco

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Por

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para obtener el grado de **Maestro (a) en Ciencias en Recursos Naturales y Desarrollo Rural**

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Resumen

Diversas especies de insectos han desarrollado resistencia o tolerancia a insecticidas. Aunque la presencia de mutaciones en genes relacionados con la degradación o insensibilidad a insecticidas se encuentra bien demostrada se han descubierto microorganismos endosimbiontes capaces de neutralizar la toxicidad de dichos compuestos, proporcionando resistencia a sus hospederos. En *Apis mellifera*, se han reportado microorganismos benéficos que aparentemente están involucrados en la degradación de pesticidas. El objetivo del trabajo fue determinar si la microbiota intestinal de *A. mellifera* y *Scaptotrigona mexicana* presentes en el Soconusco, Chiapas, es capaz de degradar al insecticida malatión debido al uso constante en campo; además se consideró al plaguicida spinetoram como un testigo de reciente utilización en campo al que no han desarrollado tolerancia. Se llevó a cabo el siguiente procedimiento experimental: 1) se realizaron colectas de abejas en apiarios y meliponarios en zonas agrícolas del Soconusco, expuestos continuamente a insecticidas; 2) se determinó la presencia de malatión en cuerpos de abejas mediante extracciones por el método de QuEChERS y se analizaron mediante cromatografía de gases; 3) se calculó la DL50 y CL50 de ambas especies a ambos insecticidas; 4) se disectaron los tractos digestivos de abejas, para evaluar los consorcios bacterianos en medio M9, enriquecido con ambos plaguicidas a diversas concentraciones. No se encontraron residuos de malatión en los especímenes analizados. La DL50 y la CL50 obtenida fue parecida a los obtenidos en estudios previos, indicando que las abejas no habían desarrollado tolerancia al malatión ni al spinetoram. No se obtuvieron consorcios bacterianos del tracto digestivo de abejas capaces de degradar malatión ni spinetoram.

Palabras clave: interacción insecto – microorganismos, endosimbiontes, tolerancia, pesticidas, exposición prolongada.

Introducción

Las abejas son los principales polinizadores de plantas con flores en el mundo; polinizan alrededor de 73% de los vegetales que consumimos y el 75% de la vegetación silvestre (Diodato et al., 2008). Dentro de los cultivos de interés económico para la región del Soconusco que son polinizados por abejas se encuentran el rambután (*Nephelium lappaceum*), el café (*Coffea* spp.), el marañón (*Anacardium occidentale*) y en cierta medida el mango (*Mangifera indica*). Sin embargo, se ha registrado un decremento de las poblaciones en abejas asociado al aumento irracional de plaguicidas (Arizmendi, 2009; Schierow et al., 2012). El insecticida dicloro difenil tricloroetano (DDT), molécula organoclorada que tiene alta estabilidad y persistencia en el ambiente, fue usada ampliamente para el control de insectos desde inicios del siglo XX, pero debido a los daños que ocasionó a la salud de los mamíferos, fue prohibido su uso en la década de los setenta, lo que llevó a la búsqueda de plaguicidas menos agresivos. En la actualidad, los insecticidas más utilizados son aquellos provenientes de las familias de neonicotinoides, piretroides, macrólidos, y en mayor proporción los organofosforados, como el malatión, de amplio uso en la agricultura y que presenta una alta afinidad por tejido lipídico, acumulándose en órganos y tejidos de importancia vital para la salud de los mamíferos (Hazarika et al., 2003; Ruiz-Toledo, 2019). Se ha demostrado que los plaguicidas compuestos por moléculas organofosforadas actúan en la fosforilación del grupo hidroxilo de la serina del sitio activo de la acetilcolinesterasa (AChE) que se encuentra en las terminaciones nerviosas, lo que conlleva a la inactivación de esta enzima esencial, que tiene un papel importante en la neurotransmisión del sistema nervioso en insectos (Sudakin & Stone, 2011; Du et al., 2012). Estudios con organofosforados como malatión y clorpirifos en *Apis mellifera* han demostrado los daños que ocasionan en insectos, destacando problemas como la pérdida de visión, habilidad de vuelo, y desorientación al momento de forrajear. Estos efectos han llevado al colapso de colmenas y a la pérdida de aproximadamente un tercio de la riqueza de especies de abejas en Norteamérica (Kluser & Peduzzi, 2007; Oliveira et al., 2014). Actualmente, han surgido nuevas alternativas que ostentan ser biodegradables y

por tanto “amigables con el ambiente”. Sin embargo, a la fecha se desconoce el daño que pueda causar en la salud de los insectos benéficos. Uno de estos insecticidas compuesto por macrólidos es el spinetoram, insecticida semisintético que está compuesto por dos moléculas, espinosinas J y L, las cuales actúan sobre el sistema neurotransmisor de los insectos de manera más rápida que su antecesor el insecticida natural Spinosad. Las espinosinas, al momento de entrar en contacto o ser ingeridas por los insectos, interrumpen los receptores nicotínicos de acetilcolina y ácido γ -aminobutírico (GABA), provocando que su sistema neurotransmisor falle, teniendo como resultado daños letales o subletales en la vida del insecto (Hussain et al., 2018). Por consiguiente, ha crecido el interés por el estudio sobre las afectaciones de los pesticidas en polinizadores, así como la búsqueda de soluciones hacia estos efectos. No obstante, se ha demostrado el potencial de algunos insectos por tener resistencia o tolerancia hacia ciertos insecticidas. Algunos estudios han demostrado que ciertos insectos tienen abundante producción de enzimas como glutatión-S-transferasas (GST), monooxigenasas del citocromo P450 (P450s) y las carboxilo / colinesterasas (CCE) que ayudan en la degradación de los plaguicidas, provocando que estos insectos presenten resistencia y tolerancia a algunos plaguicidas (Yu et al., 1984; Claudianos et al., 2006). En moscas de *Ceratitis capitata*, se ha demostrado la producción de acetilcolinesterasa (AChE) mutante, que provoca la resistencia en malatión (Magaña et al., 2008). Otra de las formas en la que los insectos han generado tolerancia hacia los insecticidas se basa en interacciones con sus microorganismos simbiotes, que permiten la asimilación y transformación de compuestos dañinos a metabolitos digeribles y de menor impacto para las abejas (Engel & Moran, 2013). Algunos estudios han demostrado el potencial de las bacterias para degradar plaguicidas como el malatión; se han descrito especies como *Pseudomonas* sp., *Pseudomonas putida*, *Micrococcus lylae*, *Pseudomona saureofaciens* y *Acetobacter liquefaciens* que pueden degradar plaguicidas (Goda et al., 2010). En cucarachas se ha descubierto una simbiosis benéfica entre bacterias, donde los microorganismos *Pseudomonas aeruginosa*, *Streptomona maltophilia*, *Bacillus atrophaeus* y *Citrobacter amolonaticus* que forman parte de su

microbiota, degradaron endosulfán hasta en 85% (Ozidal et al., 2016). Por lo anterior, las abejas podrían beneficiarse de la microbiota capaz de degradar agentes agroquímicos, lo que implicaría un mecanismo indirecto de tolerancia no estudiado hasta la fecha en abejas. Ante el desconocimiento de la capacidad de degradación de insecticidas como malatión y spinetoram por la microbiota intestinal de abejas, se evaluó la capacidad de la microbiota intestinal de *A. mellifera* y *S. mexicana* para degradar estos plaguicidas.

1 **Microbial gut consortia from two highly social bee species is not involved on the**
2 **degradation of the pesticides malathion and spinetoram**

3

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25 **Abstract**

26 In many bee colonies around the world honey and pollen reserves are contaminated
27 with residues of pesticides. Such chronic exposure might turn gut microorganisms capable
28 to degrade pesticides. In this way, bees would become tolerant to insecticides, as shown in
29 other insects. We tested this possibility in two bee species: the honey bee *Apis mellifera*
30 and the native stingless bee *Scaptotrigona mexicana*. We evaluated the insecticides
31 malathion, intensively and extensively used in the study region and spinetoram, a
32 semisynthetic insecticide derived from spinosyns, which was recently introduced but it is
33 scarcely used. We found that *A. mellifera* had larger LD50 to malathion and slightly smaller
34 LD50 to spinetoram than *S. mexicana*. In previous works *A. mellifera* showed larger LD50
35 to both insecticides than in our study; unfortunately, no previous data for *S. mexicana* was
36 available. Cultured guts did not grow bacteria in medium supplemented with malathion or
37 spinetoram. We found no evidence that the studied species had developed any tolerance
38 trait to any of the tested pesticides, nor the presence of bacteria capable of degrading them.

39

40 Keywords: LD50, LC50, *Apis mellifera*, *Scaptotrigona mexicana*, tolerance

41

42 **Resumen**

43 Se han reportado residuos de plaguicidas de polen y miel de colonias de abejas
44 sociales en todo el mundo. Esta exposición crónica podría funcionar como un agente de
45 selección de microorganismos simbioses de las abejas, que desarrollarían la capacidad de
46 degradar estas sustancias; de esta forma se incrementaría la tolerancia de las abejas hacia
47 los plaguicidas, como se ha mostrado en otros insectos. El objetivo de este trabajo fue
48 investigar esta posibilidad en dos especies de abejas: la abeja melífera *Apis mellifera* y la
49 abeja sin aguijón *Scaptotrigona mexicana*. Se evaluaron los insecticidas malatión, el cual se
50 usa intensiva y extensivamente en la región de estudio, y spinetoram, insecticida
51 semisintético derivado de spinosyn, el cual es de reciente introducción y se ha utilizado
52 muy poco. Se encontró que *A. mellifera* tuvo una LD50 a malatión mayor que *S. mexicana*,
53 y una LD50 a spinetoram ligeramente menor. Trabajos previos en *A. mellifera* reportan
54 LD50 mayores que los encontrados en el presente estudio para ambos insecticidas;
55 desafortunadamente no se tienen datos de *S. mexicana* para comparar. El cultivo de los
56 intestinos de las abejas en medios suplementados con malatión o spinetoram no presentó
57 crecimiento bacteriano. Así, no encontramos evidencia de que las especies que estudiamos
58 hayan desarrollado tolerancia alguna a malatión o a spinetoram, ni que tengan bacterias que
59 degraden dichos insecticidas.

60

61 **Keywords:** DL50, CL50, *Apis mellifera*, *Scaptotrigona mexicana*, tolerancia

62 **Introduction**

63 Bees provide us the fundamental ecological service of pollination, which enhances
64 crop production (Michener, 2000; Klein et al., 2003). However, control of arthropod pests,
65 which often involves the use of synthetic pesticides, threatens the populations of bees
66 (McLaughlin & Mineau, 1995). Since the discovery of the insecticidal properties of the
67 DDT, the variety and amount of synthetic chemicals used for crop protection has increased,
68 causing a steady parallel decrease in the population of bees (Oberemok et al., 2015). For
69 that reason it is not uncommon to find pesticides in pollen and honey in apiaries around the
70 world (Ruiz-Toledo et al., 2018).

71 Acute exposure to lethal dose of pesticides causes the dead of bees (Johnson, 2015).
72 Sublethal exposure contributes to the deterioration of organs such as the brain and midgut,
73 and impairs vision, flying ability, and causes disorientation and reduction in longevity
74 (Desneux et al., 2007), which eventually would cause the failure of the colony. However,
75 bees somehow tolerate sublethal exposure to insecticides, otherwise it would be impossible
76 to explain the presence of pesticides in pollen and honey in living colonies (Sanchez-Bayo
77 & Goka, 2014). Such tolerance can be partially explained by the presence of detoxifying
78 enzymes (Berenbaum & Johnson, 2015; Rand et al., 2015), though some studies seem to
79 indicate that their role is less relevant in honey bees than in other insects (Yu et al., 1984;
80 Claudianos et al., 2006). Other potential explanation can be found in the degrading role
81 microbiota in the intestine of bees, with which have symbiotic relationships (Martinson et
82 al., 2011). It is thought that damaging the gut bacteria community might have detrimental
83 effects on bee health, since some of them are associated with the assimilation of nutritive
84 substances (Vásquez et al., 2012). In fact, it has been shown in some insect species that gut
85 symbionts can also degrade pesticides. For instance, *Pseudomonas aeruginosa*,

86 *Streptomonas maltophilia*, *Bacillus atrophaeus* and *Citrobacter amolonaticus*, from the gut
87 of cockroaches chronically exposed to insecticides, were able to degrade endosulfan (Ozidal
88 et al., 2016). Other examples of symbiont-mediated insecticide resistance include
89 *Burkholderia* sp which turns the stinkbug *Riptortus pedestris* resistant to fenitrothion
90 (Kikuchi et al., 2012), the trichlorphon-degrading *Citrobacter* bacteria in the gut of the
91 oriental fruit fly *Bactrocera dorsalis* (Cheng et al., 2017) and *Spodoptera frugiperda* with
92 bacteria able to degrade lambda-cyhalothrin, deltamethrin, chlorpyrifos ethyl, spinosad and
93 lufenuron (Almeida et al., 2017).

94 The presence of pesticides in stored pollen and honey (Ruiz-Toledo et al., 2018)
95 invariably means that gut microbiota is exposed to such substances. Thus, it might be
96 possible that some bacteria had developed the potential to degrade insecticides. Bees could
97 benefit from such microbiota, because they would represent an indirect mechanism of
98 tolerance not studied to date in bees. Therefore the aim of this study was to investigate the
99 presence of intestinal bacteria, from the honey bee *Apis mellifera* and the stingless bee
100 *Scaptotrigona mexicana*, able to degrade the spynosin spinetoram and the
101 organophosphorous malathion. We chose these species because they are widely managed in
102 the study region and we have been keeping colonies in conventional agricultural landscapes
103 for over 20 years. Malathion is widely used for the control of several insect pests in several
104 crops, while spinetoram is recently and scarcely used, thus representing a no-exposure
105 pesticide for our bees. We expect that bees had developed some resistance to malathion but
106 no to spinetoram, due to the presence of malathion-degrading bacteria.

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110 **Materials and methods**

111 Our samplings were carried out in February - April 2018. *Apis mellifera* workers
112 were collected from three managed colonies located in our apiary in Tapachula, Chiapas,
113 Mexico. *Scaptotrigona mexicana* workers were collected from three colonies from our
114 meliponary located in Tuxtla Chico, Chiapas. Workers were subjected to any of the
115 following procedures.

116 We investigated the degree of exposure to malathion by extracting the lipophilic
117 substances of a macerate consisting of 25 foragers from each colony in one 13 x 100 glass
118 tube, following the QUEChERS method by Mullin et al. (2010). After centrifugation at
119 10,000 rpm during 10 min, the liquid phase was separated into another tube and then
120 analyzed by gas chromatography according to the method described by Ruiz-Toledo et al.
121 (2018), using the same equipment, run conditions and materials, but using an analytical
122 standard of malathion to identify the target pesticide (Sigma-Aldrich, Mexico). We did not
123 carry out the identification of spinetoram since it was not reported by local growers to be
124 used for pest control in our agricultural landscape at least in the last year.

125 In order to estimate the oral LC50 of commercial formulations of malathion and
126 spinetoram we exposed 3-5 day-old workers of both species to several concentrations of
127 malathion and spinetoram as follows: brood combs (with approximately 400 bees in the
128 larval stage) from each colony were taken. Next they were placed in a plastic container of
129 approximately 3 L with a 5 mL vial with 2M sucrose solution to feed the bees. They were
130 maintained at a temperature of 29 °C and 70% relative humidity and 12:12 light-darkness
131 cycle. Every day we transferred the emerged bees to 100 ml plastic recipients to know
132 exactly their age. Once they reached the age of exposure, were left starving for 24 h. Thirty
133 bees from each colony were orally exposed to 10, 5, 2.5, 1, 0.1 and 0 ppm of spinetoram,

134 and 30, 20, 10, 5, 2.5, 2, 1 and 0 ppm malathion in 2M sucrose solution. We registered
135 mortality 24 hours after exposure. We considered that a bee was dead when it did not
136 respond to a gentle touch with a paintbrush. In order to be able to make a comparison
137 between these species and data from literature we estimated the consumption of 2 M
138 sucrose to calculate LD50: we placed 20 workers in separated 100 ml plastic recipients and
139 one 10 μ l glass capillary filled with sucrose solution was introduced. To measure the effect
140 of evaporation we set 5 more recipients as mentioned before but without any worker inside.

141 Reagents used in the experiment to determine the presence of pesticide-degrading
142 bacteria in the gut of bees (M9 minimal salts growth medium containing 33.9g/L Na_2HPO_4 ,
143 15g/L KH_2PO_4 , 5g/L NH_4Cl , 2.5g/L NaCl , SIGMA-ALDRICH, United States; 5% glucose
144 solution, Laboratorios PISA, Mexico; Palgus (Spinetoram 6%), Dow Agrosiences,
145 Mexico; Malathion 83.6%, Agroquímica Tridente S.A de C.V., Mexico) were obtained in
146 local shops. Three groups of ten workers were orally exposed to the LD0, LD5, LD25,
147 LD50 or LD100 of both insecticides, as in the previous experiment. Next, all surviving
148 workers were sacrificed by cold (5 min at -20°C), and then immediately disinfected from
149 external microorganisms using sodium hypochlorite following Disayathanoowat et al.
150 (2012). The intestines of the sacrificed workers, of the same species, LD and insecticide,
151 were dissected, pooled and homogenized in 10 ml of sterile M9 medium (11.28g/L). One
152 milliliter of the homogenate was mixed with 9 ml of any of the following treatments at
153 30°C , pH 7: 1) M9 medium alone as negative control. 2) M9 medium enriched with
154 malathion or spinetoram at 22.5 μg as the sole carbon source, 3) M9 medium at 5% glucose
155 as a positive control of bacterial growth and 4) M9 medium at 5% glucose + malathion or

156 spinetoram at 22.5 μg . We followed bacterial growth for 30 days by observing changes in
157 turbidity in the media in all the rests.

158 Malathion was identified by comparing the peaks obtained in the extracts obtained
159 by the QUEChERS method and that of previous runs with the analytical standard. LC50
160 was estimated by a log-logistic approach using the package drc v 3.0-1 (Ritz et al., 2015) in
161 the R environment v 3.5.0 (R Development Core Team, 2012). LD50 was calculated by
162 multiplying the average consumption of each bee by the LC50. Changes in turbidity in the
163 growth medium were considered as positive if a change occurred or negative in the
164 opposite case.

165

166 **Results**

167 We could not detect malathion in any of the pooled extracts, i.e. probably foragers
168 did not carry in their body this pesticide back to the colony. Table 1 shows the LC50 and
169 95% confidence intervals for malathion and spinetoram in both species. *Scaptotrigona*
170 *mexicana* seems to be far more sensitive to malathion than *A. mellifera*. *Scaptotrigona*
171 *mexicana* consumed 2.6 μl / 24 h on average and *A. mellifera* 50 μl / 24 h. From these data
172 it could be possible to estimate LD50 (Table 1).

173 Results from the experiment about pesticide-growing bacteria are shown in Table
174 2a, b. Control treatments T1 and T3 presented turbidity as expected: no turbidity and
175 turbidity, respectively. No turbidity was observed in any of the tubes from treatment 2, so
176 microbiota was unable to use directly any of the chemicals as the only source of carbon in
177 the conditions of our experiments. Interestingly enough, in the case of *S. mexicana* some

178 tubes of T4 did not show turbidity, indicating some degree of inhibition by the pesticides.

179 But, some form *S. mexicana* and all tubes of T4 in *A. mellifera* did show bacterial growth.

180

181 **Discussion**

182 We did not detect malathion in any worker, nor observed malathion/spinetoram-
183 degrading bacteria, despite the beehives were located in an agricultural landscape where
184 producers still use malathion to control insect pests. The lack of malathion/spinetoram
185 degrading bacteria does not rule out that they are in contact with these compounds or that
186 they cannot actually degrade these substances. It is possible we could not detect malathion
187 due to the short half-life of malathion in honey and beeswax it is in the order of minutes, as
188 compared to other insecticides (Shimshoni et al., 2019). Actually, in a recent work Ruiz-
189 Toledo (2019) reported the presence of malathion in some pollen samples of the study
190 region, indicating recent exposure. The minute amounts of malathion brought back by
191 foragers perhaps did not represent a threat to our target species, and could be easily
192 degraded by gut enzymes, as shown by our data of LD50 in spinetoram and malathion
193 which are within the parameters described by other studies conducted in *A. mellifera*
194 (Naggar et al., 2015; Shimokawatoko et al., 2012). Also it is possible that foragers
195 receiving lethal doses die before returning to the colony, thus making it undetectable in our
196 specimens and minimizing the contact with the colonies. Finally, the presence of
197 organochlorines (OCs) in the samples analyzed by Ruiz-Toledo et al. (2018), indicates that
198 exposure to organophosphorous compound might be lower than expected, and that search
199 should go in the direction of OCs-degrading bacteria.

200 To be able to develop some pesticide-degrading capacity, microorganisms should be
201 exposed to the insecticide and survive; the insect harboring the microorganisms must also

202 survive to the exposure. And then, some means to reproduce the microorganisms beyond
203 the dead of the host must be present. In the case of highly social bees, exposure to sublethal
204 amounts of insecticides occurs worldwide (Mitchell et al., 2017; Ruiz-Toledo et al., 2018),
205 so microorganisms are also exposed, but maybe at amounts that do not represent a selective
206 factor in microbial communities; on the other hand, Tian et al. (2012) found tetracyclin-
207 resistant gut microbiota in long-term exposed honey bee colonies that were directly under
208 selective pressure. Once established, the transmission of microbiota within the hive seems
209 to be through trophallaxis by nurse bees to younger stages (Powell et al., 2014). In the case
210 of insecticides they potentially had little selective pressure over microbiota, so it was not
211 possible to detect degrading bacteria. However, soil microorganisms are more likely to
212 develop pesticide-degrading capabilities. For instance, the bean bug, *Riptortus pedestris*,
213 acquire fenitrothion-degrading microorganisms from soil during larval stage (Kikuchi et al.,
214 2012). Perhaps symbiont-conferred resistant is present in social and non-social bee species
215 that use soil as the preferred nesting substrate, so it is less likely to find them otherwise.
216 This has yet to be investigated.

217

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References

227

228 Al Naggar, Y., Codling, G., Vogt, A., Naiem, E., Mona, M., Seif, A., & Giesy, J. P. (2015).

229 Organophosphorus insecticides in honey, pollen and bees (*Apis mellifera* L.) and

230 their potential hazard to bee colonies in Egypt. *Ecotoxicology and environmental*

231 *safety*, 114, 1-8.

232 Almeida, L. G. d., Moraes, L. A. B. d., Trigo, J. R., Omoto, C., & C nsoli, F. L. (2017).

233 The gut microbiota of insecticide-resistant insects houses insecticide-degrading

234 bacteria: A potential source for biotechnological exploitation. *PLoS One*, 12,

235 e0174754.

236

237 Berenbaum, M. R., & Johnson, R. M. (2015). Xenobiotic detoxification pathways in honey

238 bees. *Current Opinion in Insect Science*, 10, 51-58.

239

240 Cheng, D., Guo, Z., Riegler, M., Xi, Z., Liang, G., & Xu, Y. (2017). Gut symbiont

241 enhances insecticide resistance in a significant pest, the oriental fruit fly *Bactrocera*

242 *dorsalis* (Hendel). *Microbiome*, 5, 13.

243

244 Claudianos, C., Ranson, H., Johnson, R., Biswas, S., Schuler, M., Berenbaum, M.,

245 Feyereisen, R., & Oakeshott, J. (2006). A deficit of detoxification enzymes:

246 pesticide sensitivity and environmental response in the honeybee. *Insect Molecular*

247 *Biology*, 15, 615-636.

248

249 Desneux, N., Decourtye, A., & Delpuech, J. M. (2007). The sublethal effects of pesticides
250 on beneficial arthropods. *Annual Review of Entomology*, 52, 81-106.

251

252 Disayathanoowat, T., Yoshiyama, M., Kimura, K., & Chantawannakul, P. (2012). Isolation
253 and characterization of bacteria from the midgut of the Asian honey bee (*Apis*
254 *cerana indica*). *Journal of Apicultural Research*, 51, 312-319.

255

256

257 Johnson, R. M. (2015). Honey bee Toxicology. *Annual Review of Entomology*, 60, 415-
258 434.

259

260 Kikuchi, Y., Hayatsu, M., Hosokawa, T., Nagayama, A., Tago, K., & Fukatsu, T. (2012).
261 Symbiont-mediated insecticide resistance. *Proceedings of the National Academy of*
262 *Sciences*, 109.

263

264 Klein, A. M., Steffan–Dewenter, I., & Tscharntke, T. (2003). Fruit set of highland coffee
265 increases with the diversity of pollinating bees. *Proceedings of the Royal Society of*
266 *London. Series B: Biological Sciences*, 270, 955-961.

267

268 Martinson, V. G., Danforth, B. N., Minckley, R. L., Rueppell, O., Tingek, S., & Moran, N.
269 A. (2011). A simple and distinctive microbiota associated with honey bees and
270 bumble bees. *Molecular Ecology*, 20, 619-628.

271

- 272 McLaughlin, A., & Mineau, P. (1995). The impact of agricultural practices on biodiversity.
273 *Agriculture, Ecosystems & Environment*, 55, 201-212.
274
- 275 **Michener, C. D. 2000.** The bees of the world, vol. 1, JHU press.
276
- 277 Mitchell, E. A. D., Mulhauser, B., Mulot, M., Mutabazi, A., Glauser, G., & Aebi, A.
278 (2017). A worldwide survey of neonicotinoids in honey. *Science*, 358, 109.
279
- 280 Mullin, C. A., Frazier, M., Frazier, J. L., Ashcraft, S., Simonds, R., van Engelsdorp, D., &
281 Pettis, J. S. (2010). High levels of miticides and agrochemicals in North American
282 apiaries: Implications for honey bee health. *PLoS One*, 5, e9754.
283
- 284 Oberemok, V. V., Laikova, K. V., Gninenko, Y. I., Zaitsev, A. S., Nyadar, P. M., &
285 Adeyemi, T. A. (2015). A short history of insecticides. *Journal of Plant Protection*
286 *Research*, 55, 221-226.
287
- 288 Powell, J. E., Martinson, V. G., Urban-Mead, K., & Moran, N. A. (2014). Routes of
289 acquisition of the gut microbiota of the honey bee *Apis mellifera*. *Applied and*
290 *Environmental Microbiology*, 80, 7378.
291
- 292 **R Development Core Team 2012.** R: A language and environment for statistical
293 computing computer program, version 3.5.0. By R Development Core Team,
294 Vienna, Austria.
295

- 296 Rand, E. E. d., Smit, S., Beukes, M., Apostolides, Z., Pirk, C. W. W., & Nicolson, S. W.
297 (2015). Detoxification mechanisms of honey bees (*Apis mellifera*) resulting in
298 tolerance of dietary nicotine. *Scientific Reports*, 5, 11779.
- 299
- 300 Ritz, C., Baty, F., Streibig, J. C., & Gerhard, D. (2015). Dose-response analysis using R.
301 *PLoS One*, 10, e0146021.
- 302
- 303 **Ruiz-Toledo, J. 2019.** Efecto de agroquímicos utilizados en cultivos tropicales sobre
304 polinizadores. PhD, El Colegio de la Frontera Sur (ECOSUR) Tapachula, Chiapas.
- 305
- 306 Ruiz-Toledo, J., Vandame, R., Castro-Chan, R. A., Penilla-Navarro, R. P., Gómez, J., &
307 Sánchez, D. (2018). Organochlorine pesticides in honey and pollen samples from
308 managed colonies of the honey bee *Apis mellifera* Linnaeus and the stingless bee
309 *Scaptotrigona mexicana* Guérin from Southern, Mexico. *Insects*, 9, 54.
- 310
- 311 Sanchez-Bayo, F., & Goka, K. (2014). Pesticide residues and bees – A risk assessment.
312 *PLoS One*, 9, e94482.
- 313
- 314 Shimokawatoko, Y., Sato, N., Yamaguchi, Y., & Tanaka, H. (2012). Development of the
315 novel insecticide spinetoram (Diana®). *Sumitomo Chemical Co., Ltd., Tokyo*.
- 316
- 317 Shimshoni, J. A., Sperling, R., Massarwa, M., Chen, Y., Bommuraj, V., Borisover, M., &
318 Barel, S. (2019). Pesticide distribution and depletion kinetic determination in honey

- 319 and beeswax: Model for pesticide occurrence and distribution in beehive products.
320 *PLoS One*, 14, e0212631.
- 321
- 322 Tian, B., Fadhil, N. H., Powell, J. E., Kwong, W. K., & Moran, N. A. (2012). Long-term
323 exposure to antibiotics has caused accumulation of resistance determinants in the
324 gut microbiota of honeybees. *mBio*, 3, e00377-00312.
- 325
- 326 Vásquez, A., Forsgren, E., Fries, I., Paxton, R. J., Flaberg, E., Szekely, L., & Olofsson, T.
327 C. (2012). Symbionts as major modulators of insect health: lactic acid bacteria and
328 honeybees. *PLoS One*, 7, e33188.
- 329
- 330 Yu, S. J., Robinson, F. A., & Nation, J. L. (1984). Detoxication capacity in the honey bee,
331 *Apis mellifera* L. *Pesticide Biochemistry and Physiology*, 22, 360-368.

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333 List of tables

334 Table 1. LC50 in ppm (LD in ng/bee) of spinetoram and malathion in the tested species

335 with showing 95% confidence interval.

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337 Table 2a and 2b. Results of the experiments to detect insecticide-degrading bacteria in

338 *Scaptotrigona mexicana* and *Apis mellifera*. LD column refers to the percentil of the lethal

339 dose estimate in the corresponding experiment . T1: M9 medium alone as negative control.

340 T2: M9 medium enriched with the insecticide at 22.5 µg, as the sole carbon source, T3: M9

341 medium at 5% glucose as a positive control of bacterial growth and T4: M9 medium at 5%

342 glucose + insecticide at 22.5 µg. Turbidity is expressed as either positive (bacterial growth)

343 or negative (no growth).

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345 Table 1.

Spinetoram (ppm)					
	LC50 (LD50)	Lower Limit	Upper Limit	Best Fit Model	LD50 in other studies
<i>S. mexicana</i>	4.5 ppm (11.7 ng/bee)	2.3	6.6	Weibull	No study available
<i>A. mellifera</i>	0.21 ppm (10.5 ng/bee)	0.03	0.38	Log-logistic	24.8 ng/bee (Shimokawatoko et al., 2012)
Malathion (ppm)					
	LC50 (LD50)	Lower Limit	Upper Limit	Best Fit Model	LD50 in other studies
<i>S. mexicana</i>	10.0 ppm (26 ng/bee)	7.8	12.3	Log-logistic	No study available
<i>A. mellifera</i>	3.6 ppm (180 ng/bee)	2.0	5.15	Weibull	335 ng/bee (Al Naggari et al., 2015)

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349 Table 2a.

<i>Scaptotrigona mexicana</i>		TREATMENT			
Insecticide	LD	T1	T2	T3	T4
Malathion	0	Negative	Negative	Positive	Negative
Malathion	5	Negative	Negative	Positive	Negative
Malathion	25	Negative	Negative	Positive	Negative
Malathion	50	Negative	Negative	Positive	Positive
Malathion	100	Negative	Negative	Positive	Positive
Spinetoram	0	Negative	Negative	Positive	Positive
Spinetoram	5	Negative	Negative	Positive	Negative
Spinetoram	25	Negative	Negative	Positive	Positive
Spinetoram	50	Negative	Negative	Positive	Negative
Spinetoram	100	Negative	Negative	Positive	Negative

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363 Table 2b.

<i>Apis mellifera</i>		TREATMENT			
Insecticide	LD	T1	T2	T3	T4
Malathion	0	Negative	Negative	Positive	Positive
Malathion	5	Negative	Negative	Positive	Positive
Malathion	25	Negative	Negative	Positive	Positive
Malathion	50	Negative	Negative	Positive	Positive
Malathion	100	Negative	Negative	Positive	Positive
Spinetoram	0	Negative	Negative	Negative	Positive
Spinetoram	5	Negative	Negative	Positive	Positive
Spinetoram	25	Negative	Negative	Positive	Positive
Spinetoram	50	Negative	Negative	Positive	Positive
Spinetoram	100	Negative	Negative	Positive	Positive

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Conclusiones

Con las condiciones empleadas en este trabajo, no se detectaron microorganismos capaces de degradar malatión, concordando con los datos obtenidos en cromatografía de gases en donde tampoco se logró detectar la presencia de malatión en las abejas *Apis mellifera* colectadas en campo. Es poco probable que se encuentren abejas expuestas a dosis sub-letales, dado que realizamos las colectas durante los periodos de mayor aspersión de malatión y spinetoram en los sitios de muestreo.

Se encontró que *A. mellifera* tuvo una LD50 a malatión mayor que *S. mexicana*, y una LD50 a spinetoram ligeramente menor.

Las colonias utilizadas durante los experimentos realizados no presentaron daños aparentes por las prácticas agrícolas realizadas en la zona.

Faltan pruebas a realizar entre especies solitarias y abejas sociales para poder comprender sí la susceptibilidad de las abejas está ligada a la procedencia de su microbiota intestinal.

Literatura citada

- Arizmendi C. 2009. La crisis de los polinizadores. *Biodiversitas*. 85:1-5.
- Claudianos C, Ranson H, Johnson R, Biswas S, Schuler M, Berenbaum M, Feyereisen R, Oakeshott J. 2006. A deficit of detoxification enzymes: Pesticide sensitivity and environmental response in the honeybee. *Insect molecular biology* 15(5):615-636.
- Diodato L, Fuster A, Maldonado M. 2008. Valor y beneficios de las abejas nativas, (Hymenoptera: Apoidea), en los bosques del chaco semiárido, Argentina. *Quebracho-Revista de Ciencias Forestales*. (15).
- Du D, Wang J, Wang L, Lu D, Lin Y. 2012. Integrated lateral flow test strip with electrochemical sensor for quantification of phosphorylated cholinesterase: Biomarker of exposure to organophosphorus agents. *Analytical Chemistry*. 84(3):1380-1385.
- Engel P, Moran NA. 2013. The gut microbiota of insects—diversity in structure and function. *FEMS Microbiol Reviews*. 37(5): 99-735.
- Goda, S. K., Elsayed, I. E., Khodair, T. A., El-Sayed, W., & Mohamed, M. E. (2010). Screening for and isolation and identification of malathion-degrading bacteria: cloning and sequencing a gene that potentially encodes the malathion-degrading enzyme, carboxylestrase in soil bacteria. *Biodegradation*. 21(6): 903-913.
- Hazarika A, Sarkar SN, Hajare S, Kataria M, Malik JK. 2003. Influence of malathion pretreatment on the toxicity of anilofos in male rats: A biochemical interaction study. *Toxicology*. 185(1-2):1-8.

- Hussain A, Khan MF, Mukhtar O, Faheem M, Ali U. 2018. Comparative toxicity of spinetoram and nitenpyram against earthworm and their effects on protein contents and cholinesterase activity. *Biological Sciences-PJSIR*. 61(2):103-111.
- Kluser S, Peduzzi P. 2007. Global pollinator decline: A literature review. UNEP/GRID-Europe.
- Magaña C, Hernández-Crespo P, Brun-Barale A, Couso-Ferrer F, Bridce JM, Castañera P, Feyereisen R, Ortego F. 2008. Mechanisms of resistance to malathion in the medfly *Ceratitis capitata*. *Insect Biochemistry and Molecular Biology*. 38(8):756-762.
- Oliveira RA, Roat TC, Carvalho SM, Malaspina O. 2014. Side-effects of thiamethoxam on the brain and midgut of the africanized honey bee *Apis mellifera* (Hymenoptera: Apidae). *Environmental Toxicology*. 29(10):1122-1133.
- Ozdal M, Ozdal OG, Algur OF. 2016. Isolation and characterization of α -endosulfan degrading bacteria from the microflora of cockroaches. *Polish Journal of Microbiology*. 65(1):63-68.
- Ruiz-Toledo J. 2019. Efecto de agroquímicos utilizados en cultivos tropicales sobre polinizadores. Tapachula, Chiapas. [Tesis de Doctorado] El Colegio de la Frontera Sur (ECOSUR). 7-10.
- Schierow LJ, Johnson R, Corn ML. 2012. Bee health: The role of pesticides. Congressional Research Service.7-57.

Sudakin DL, Stone DL. 2011. Dialkyl phosphates as biomarkers of organophosphates: The current divide between epidemiology and clinical toxicology. *Clinical Toxicology*. 49(9):771-781.

Yu SJ, Robinson FA, Nation JL. 1984. Detoxication capacity in the honey bee, *Apis mellifera* L. *Pesticide Biochemistry and Physiology*. 22(3):360-368.